

1     **The scope of antimicrobial resistance in Residential Aged Care Facilities determined**  
2             **through analysis of *Escherichia coli* and the total wastewater resistome**

3     **Sylvia A. Sapula<sup>1</sup>, Anteneh Amsalu<sup>1,2</sup> Jon J. Whittall<sup>1</sup>, Bradley J. Hart<sup>1</sup>, Naomi L.**  
4     **Siderius<sup>1</sup>, Lynn Nguyen, Cobus Gerber, John Turnidge<sup>3</sup>, Henrietta Venter<sup>1\*</sup>**

5             1. UniSA Clinical and Health Sciences, Health and Biomedical Innovation, University  
6             of South Australia, Adelaide, SA 5000, Australia

7             2. Department of Medical Microbiology, University of Gondar, Gondar 196, Ethiopia

8             3. Adelaide Medical School and School of Biological Sciences University of Adelaide,  
9             Adelaide, SA 5000, Australia

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11             \*Correspondence: rietie.venter@unisa.edu.au; Tel.: +61-8-8302-1515

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17     **ABSTRACT**

18     High and often inappropriate antibiotic use has been documented for residential aged care  
19     facilities (RACFs). As a result, RACFs represent selective environments favoring the  
20     development of antimicrobial resistance (AMR). However, surveillance of the prevalence and  
21     scope of AMR in RACFs is limited. Here, wastewater-based epidemiology encompassing  
22     culture-based methods in combination with whole genome sequencing and metagenomics  
23     allowed for the in-depth analysis of the prevalence of antimicrobial-resistant *Escherichia coli*

24 in two RACFs and one Retirement facility. Wastewater was collected at five different time  
25 points over 18 months. From these, *E. coli* were isolated and assessed for phenotypic and  
26 genotypic resistance. The antimicrobial resistome of each wastewater sample was also  
27 determined. A comparison of facilities revealed a higher prevalence of AMR and multidrug-  
28 resistant *E. coli* observed in one RACF. The international high-risk *E. coli* clone, ST131,  
29 carrying CTX-M-like extended-spectrum beta-lactamases was exclusively isolated from this  
30 facility. A high proportion of these isolates were classified as belonging to phylogroups B2  
31 and D, associated with virulent extra-intestinal *E. coli* infections. The resistome of this  
32 facility also revealed a higher prevalence of mobile resistance genes such as *sul1* and *sul2*,  
33 conferring sulfamethoxazole resistance, the plasmid-mediated quinolone-resistance (PMRQ)  
34 gene *qnrS*, and *qacEdelta1*, that confers biocide resistance. High fluoroquinolone resistance  
35 rates were observed in all three sample sites despite measures in place limiting the use of this  
36 class of antibiotics. The findings of this study illustrate that RACFs are highly selective  
37 environments that require measures to limit AMR development, potentially through  
38 antimicrobial stewardship.

39

#### 40 **IMPORTANCE**

41 Antimicrobial resistance (AMR) is a global threat that imposes a heavy burden on our health  
42 and economy. Residential aged care facilities (RACFs), where frequent inappropriate  
43 antibiotic use creates a selective environment that promotes the development of bacterial  
44 resistance significantly contribute to this problem. We used wastewater-based epidemiology  
45 to provide a holistic whole-facility assessment and comparison of antimicrobial resistance in  
46 two RACFs and a retirement village. Resistant *E. coli*, a common and oftentimes problematic  
47 pathogen within RACFs was isolated from the wastewater and the phenotypic and genotypic

48 AMR was determined for all isolates. We observed a high prevalence of an international  
49 high-risk clone, carrying an extended-spectrum beta-lactamase in one facility. Analysis of the  
50 entire resistome also revealed a greater number of mobile resistance genes in this facility.  
51 Finally, both facilities displayed high fluoroquinolone resistance rates – a worrying trend  
52 seen globally despite measures in place aimed at limiting their use.

53

54 **KEYWORDS** Antimicrobial resistance, wastewater-based epidemiology, *Escherichia coli*,  
55 multidrug resistance, metagenomics

56

## 57 **INTRODUCTION**

58 Antimicrobial resistance (AMR) is an urgent and global threat [1], with recent reports  
59 estimating that approximately 4.95 million AMR associated deaths occurred in 2019 [2].  
60 Continual AMR development is also speculated to add to the global economic burden, as  
61 projections indicate that 300 million people will lose their lives to AMR over the next 35  
62 years, resulting with a loss of 60 to 100 trillion USD worth of economic output [1]. In  
63 addition to increased mortality rates, AMR can lead to increased duration of illness  
64 particularly for those who are immunocompromised [3, 4]. In the elderly population, where  
65 individuals are at an increased risk of incidence and severity of bacterial infections, AMR  
66 bacteria contribute significantly to the burden of infection, with infectious diseases  
67 accounting for one third of all deaths in those aged 65 years and over [5-7].

68 A significant contributor to the development and spread of AMR is the overuse and misuse of  
69 antibiotics, which promotes the selection of AMR bacteria [8-11]. Residential aged care  
70 facilities (RACFs) exemplify the overuse and frequent inappropriate use of antibiotics [12-  
71 14]. This coupled with the vulnerability of the residents resulting from comorbidities, a

72 compromised immune system, close living proximities, hospital visits and frequent healthcare  
73 worker contact, facilitate the spread of resistant bacteria and promote a high infection burden  
74 among the residents [15-17]. As such, RACFs and residents themselves have been identified  
75 as important reservoirs for the development of emerging AMR and multidrug resistant  
76 (MDR) bacteria [18].

77 Surveillance has often been limited to healthcare setting such as hospitals, where pathogens  
78 isolated from clinical infections are reported [19, 20]. Surveillance studies of these in  
79 wastewater from RACFs are scant [21]. Studies assessing AMR in RACFs are resident  
80 focussed and, although vital, are fraught with limitations regarding the number of residents  
81 included in each study, and as such, do not always represent the facility as a whole.  
82 Nonetheless, such studies have reported a high prevalence of MDR extended-spectrum beta-  
83 lactamase (ESBL) producing *Escherichia coli* [22, 23]. A point prevalence survey of three  
84 Melbourne based RACFs revealed that 27 % of residents were colonised with ESBL-  
85 producing *E. coli*, which were clonal in nature [22]. An increasing prevalence of ESBL-  
86 producing *E. coli* has also been observed in aged care facilities in Germany [24, 25], Japan  
87 [26] and the Netherlands [27]. ESBL-producing ST131 *E. coli* were also observed in aged  
88 care facilities in Germany [24] and Portugal [28]. This clonal type, also referred to as a  
89 worldwide pandemic clone [29], has spread rapidly throughout different parts of the world  
90 since its discovery in 2008 [30]. In addition to ESBL expressing ST131, a UK study has also  
91 found fluoroquinolone-resistant but ESBL lacking ST131 strains carrying the plasmid-  
92 mediated *aac(6)-Ib-cr* gene, which reduced susceptibility to ciprofloxacin [31], suggesting a  
93 split and continual emergence of ST131. In Australia, rates of *E. coli* found to be resistant to  
94 beta-lactams such as ceftriaxone and fluoroquinolones such as ciprofloxacin continue to  
95 increase, despite restrictions being in place to limit access to these agents [32]. The

96 prevalence of these underscores the importance of surveillance studies in RACF, with the use  
97 of wastewater-based surveillance offering an insight into the facility as whole.

98 Therefore, the aim of this study was to utilise wastewater-based epidemiology, consisting of  
99 microbial culturing, whole genome sequencing and metagenomics to investigate AMR within  
100 two Adelaide RACFs. The use of wastewater in this thorough investigation has circumvented  
101 one of the most common limitations in studies of AMR in RACFs - low recruitment numbers  
102 of participants - and allowed for a holistic approach to the assessment of AMR in entire  
103 RACFs. This study demonstrates the importance of surveillance studies focussing on AMR  
104 within RACFs. Housing a vulnerable older population, the development of bacterial  
105 resistance within these facilities contributes to a growing burden of infection for which  
106 treatment may become more limited. As an ageing population, with 1 in 6 estimated to be  
107 aged over 60 years old by 2030 [33], the development of AMR within RACFs, as such, is  
108 area of great concern. This study exemplifies a novel approach to the assessment of AMR  
109 within RACFs, as the use of wastewater, microbial culturing, whole genome sequencing and  
110 metagenomics to assess the development of microbial resistance has not previously been  
111 undertaken in this setting. The inclusion of one facility which has implemented an  
112 antimicrobial stewardship (AMS) program and another which has not, also allowed for an  
113 assessment of the effect that an AMS program may have on the prevalence of AMR bacteria  
114 within a facility.

115

## 116 **MATERIALS AND METHODS**

### 117 **Sampling**

118 Wastewater was sampled from two RACFs (Facility 1 with 170 beds and Facility 2 with 70  
119 beds) and one Retirement village (38 apartments with residents of similar age to those from

120 the RACFs) in Adelaide. All sites are managed by the same service provider and are located  
121 within a 20 km radius of each other. Of the two RACFs Facility 2 has had an AMS program  
122 (Supplemental 1) implemented for 3 years prior to this study, whilst Facility 1 does not have  
123 AMS program implemented. As both aged care facilities are managed by the same care  
124 provider, they are comparable in the amenities and care provided. Both facilities have long-  
125 term residents in addition to respite care.

126

127 Wastewater samples were collected at approximately three-month intervals at five different  
128 time points from October 2019 to February 2021. Grab samples of approximately 200 ml  
129 were collected every hour over a 10-hour period, with collection starting at 7 am to capture  
130 the morning routine. These were collected from an access point which captured all of  
131 wastewater from the RACF, before it flowed into the main sewage system. Samples were  
132 then stored on ice and pooled for analysis. Sampled wastewater was transported to the  
133 laboratory on ice, stored at 4 °C and analyzed on the day of collection.

134

135 **Isolation of *Escherichia coli***

136 AMR *E. coli* isolates were screened on selective and differential plates. For isolation of *E.*  
137 *coli* from wastewater, 100 µl wastewater (neat and 1:10 dilution) was plated onto selective  
138 and differential plates. Presumptive identification of *E. coli* was determined using two sets of  
139 Brilliance™ *E. coli*/coliform (CM0956, Oxoid, Australia) plates supplemented with either 1  
140 mg/L ceftazidime or 0.5 mg/L of ciprofloxacin and Tergitol 7 (CM0793, Oxoid) plates  
141 supplemented with 0.5 mg/L ciprofloxacin. Cultures were incubated at 37 °C for 18 h.  
142 Following colony purification, identification was verified by matrix-assisted laser  
143 desorption/ionization time-of-flight mass spectrometry (MALDI-TOF) (Bruker Daltonik  
144 GmbH, Bremen, Germany). Confirmed *E. coli* isolates were stored at -80 °C in Tryptone  
145 Soya broth (CM0129, Oxoid) supplemented with 20 % (v/v) glycerol.

146

147 **Antimicrobial susceptibility testing**

148 Antibiotics assessed in this study included cefepime, ceftazidime, piperacillin-tazobactam,  
149 gentamicin, meropenem, trimethoprim-sulfamethoxazole, colistin (ChemSupply, Australia)  
150 and ciprofloxacin (Sigma-Aldrich, Australia). The Minimum Inhibitory Concentration (MIC)  
151 of each antibiotic assessed in this study was determined using the broth dilution method as  
152 described by the ISO standard [34-36]. Briefly, a 96 well microtiter plate was prepared with  
153 twofold serial dilutions of the antimicrobial agent. Cultures at an OD<sub>600</sub> of 0.0025 nm were  
154 used as an inoculum using Mueller Hinton broth as a growth medium. *E. coli* ATCC 25922  
155 was used as a quality control strain. The 96 well microtiter plates were incubated at 37 °C  
156 with 120 rpm shaking for 16 h. Resistance/susceptibility was determined by measuring  
157 bacterial growth by optical density at 600 nm using a Thermofisher Multiskan FC photometer  
158 (Thermofisher, Australia) using *E. coli* ATCC25922 for quality control. Epidemiological  
159 cutoff values, as reported by the European Committee on Antimicrobial Susceptibility

160 Testing (<https://mic.eucast.org/search/>) were used to distinguish wild type from non-wild type  
161 isolates. In this manuscript, the terms “susceptible” and “resistant” are used to refer to “wild  
162 type” and “non-wild type”. Moreover, resistance to at least one antimicrobial agent in  $\geq 3$   
163 antimicrobial categories were considered as MDR [37]. To ensure assessment was not carried  
164 out on duplicate *E. coli* isolates, the resistance profile of all *E. coli* was evaluated. Isolates  
165 observed to have the same resistance profile and that were isolated from the same site at the  
166 same time-point were subsequently eliminated from further assessment.

167

## 168 **DNA extraction**

### 169 *Genomic extraction for whole genome sequencing*

170 Following colony purification, a single colony was incubated in LB broth and grown at 37 °C  
171 overnight. Bacterial cells were obtained by centrifugation (5000 x g, 5 min) of 5 ml of culture  
172 broth. Genomic DNA was extracted from the pellet using the MN NucleoSpin®Microbial  
173 DNA kit (Machery-Nagel GmbH and Co.KG, Duren, Germany), following manufacturer’s  
174 instructions. DNA quality and quantity were assessed using a Cytation5 imaging reader  
175 (BioTek instruments, Winoosi Vermont, USA). Extracted genomic DNA was stored at -20 °C  
176 [35, 38].

177

### 178 *Genomic extraction for shotgun metagenomic sequencing*

179 Wastewater samples (100 ml) were initially filtered through an 8 µm cellulose nitrate filter to  
180 remove debris and then through a 0.2 µm pore size cellulose nitrate filter (both from  
181 Sartorius, Goettingen, Germany). Following filtration, genomic DNA was extracted from the  
182 membranes using the DNeasy® PowerWater kit (Qiagen, Hilden, Germany), following  
183 manufacturer’s instructions. DNA extracts were stored at -20 °C prior to being analyzed by  
184 metagenomic sequencing. DNA quantity and quality were assessed spectrophotometrically



185 using the Cytation5 imaging reader (BioTek instruments, Winoosi Vermont, USA) and the  
186 Take3 MicroVolume plate.

187

### 188 *Whole genome sequencing and bioinformatic analysis*

189 Whole genome sequencing (WGS) was performed at SA Pathology (Adelaide, Australia)  
190 using the Illumia NextSeq platform. Sequencing libraries were prepared using the Nextera  
191 XT DNA library preparation kit (Illumina Inc., USA) as per manufacturer's instructions.  
192 WGS was performed on the Illumina NextSeq 550 platform with the NextSeq 500/550 Mid-  
193 Output kit v2.5 (300 cycles) (Illumina Inc). Raw 150 bp paired-end reads were used as input  
194 data for the TORMES v.1.2 [39] pipeline for the analysis of whole bacterial genomes. This  
195 included sequence quality filtering (PRINSEQ v.0.20.4) [40], *de novo* genome assembly  
196 (SPAdes v.13.4.1) [41] and annotation (Prokka v1.14.5) [42]. *E. coli* multilocus sequencing  
197 typing (MLST) profiles were predicted using mlst v2.10 (T. Seemann,  
198 <https://github.com/tseemann/mlst>) which uses the PubMLST database [43]. AMR genes were  
199 screened using ABRicate (T. Seemann, <https://github.com/tseemann/abricate>) against the  
200 ResFinder [44], Comprehensive Antibiotic Resistance Database (CARD) [45] and ARG-  
201 ANNOT [46], databases. Additional software such PointFinder [47] was used to screen for  
202 chromosomal point mutations, and PlasmidFinder [48] was used for plasmid replicon  
203 screening. Finally, FimTyper [49] and SeroTypeFinder [50] were used to type *E. coli* isolates,  
204 whilst *in silico* phylogroup determination was carried out by ClermonTyping [51]. *E. coli*  
205 AMR determinants and plasmids were further analyzed using NCBI Blast [52].

206

207 A maximum-likelihood phylogenetic tree was constructed to assess *E. coli* clonality. Core  
208 genome single nucleotide polymorphisms (SNPs) were identified using CSI phylogeny 1.4  
209 [53]. Reads from each *E. coli* isolate were aligned to a reference genome EC958 (GenBank

210 accession no. NZ\_HG941718.1). EC958 is an international MDR *E. coli* O25b:H4-ST131  
211 clone which produces the CTX-M-15 ESBL and is fluoroquinolone resistant [54]. The output  
212 file was used to annotate and visualize a phylogenetic tree using the interactive tree of life v.6  
213 (iTOL) tool [55, 56]. Unless otherwise noted, default parameters were used for all software  
214 tools.

215

### 216 **Metagenomic sequencing and bioinformation analysis**

217 A total of 15 samples were used for shotgun metagenomic sequencing and analysis.  
218 Metagenomic sequencing was performed at the SAGC (South Australian Genomics Centre,  
219 Adelaide, Australia), and Macrogen Inc. (Seoul, South Korea), using the Illumina Novaseq  
220 S1 platform with 150 bp paired ends reads. Read quality was assessed using FastQC [57].  
221 The SqueezeMeta v1.4 [58] pipeline was used for standard metagenomic analysis.  
222 SqueezeMeta uses Trimmomatic [59] for filtering and trimming of adapters before assembly  
223 using MEGAHIT [60]. The resulting contigs were filtered for quality with short contigs  
224 removed with Prinseq [40]. Gene and rRNA predication were performed utilizing the  
225 Prodigal gene prediction software [61] while rRNA sequences were found using barrnap  
226 [62]. AMR analysis was performed using SqueezeMeta v1.4 [58] by Diamond [63] against  
227 version 3.1.4 of the CARD [45]. AMR gene abundance was calculated as transcripts per  
228 million (TPM) mapped bacterial reads.

229

### 230 **Statistical analysis**

231 Principal components analysis (PCA) was performed in R studio v.1.2.5033 and used here to  
232 visually compare the distribution of resistant *E. coli* isolates recovered from wastewater  
233 sampled from different sites (facilities). Statistical analysis, bar graphs and boxplots were  
234 generated using Graph Prism v9, with statistical differences between selected AMR gene

235 prevalence and facility assessed by a two-tailed Mann-Whitney U test with results considered  
236 statistically significant at  $p < 0.05$ ;  $p < 0.01$  level.

237

#### 238 **Data availability**

239 Whole genome and shotgun metagenomic sequences were deposited in the NCBI database  
240 under Bio Project number PRJNA861152. Reads mapping to the human reference genome  
241 (GCA\_000001405.15) were removed prior to submission to public sequence databases  
242 according to the protocol used in the Human Microbiome Project [64, 65].

243

## 244 **RESULTS**

### 245 **Resistance profiles of the *E. coli* isolates reveals a high prevalence of resistance and** 246 **high-level resistance**

247 Wastewater sampling and analysis was used to evaluate the prevalence and antimicrobial  
248 resistance of *E. coli* in two local RACFs and one Retirement village. Since fluoroquinolone  
249 resistance and ESBL production is of particular concern for *E. coli*, selection of *E. coli*  
250 isolates was carried out with selective and differential media supplemented with  
251 ciprofloxacin or ceftazidime.

252 A total of 93 AMR *E. coli* isolates (n=58 from Facility 1, n=27 from Facility 2 and n=8 from  
253 a Retirement Village) were purified from wastewater samples. As expected, due to the  
254 selection process used in this study, resistance to ceftazidime and ciprofloxacin was observed  
255 for a large percentage of *E. coli* isolates, with 66.7 % (n=62/93) and 96.8 % (n=90/93) of all  
256 isolates shown to be ceftazidime and ciprofloxacin resistant respectively (Table 1).

257 Interestingly, almost all the isolates were resistant to ciprofloxacin irrespective the media and  
258 antibiotic selection/non-selection used for isolation. Also, a high incidence of trimethoprim-

259 sulfamethoxazole (n=51/93, 54.8 %) and gentamicin (n=47/93, 50.5 %) resistant *E. coli* was  
 260 also observed although resistance to these antibiotics were not selected for. Finally, resistance  
 261 to the last resort antibiotics meropenem (n=12/93, 20.7 %) and colistin (n=4/93, 4.3 %) was  
 262 observed, albeit at lower frequencies.

263

264 **Table 1.** MIC distribution for 93 *E. coli* isolates recovered from wastewater sampled from associated  
 265 RACFs and a retirement village.

Number of *E. coli* isolates with MIC (mg/L) at:

Antibiotic	0.008	0.016	0.03	0.06	0.125	0.25	0.5	1	2	4	8	16	32	64	>64	%NWT
FEP	0	0	3	19	16	5	3	10	6	1	4	6	8	1	11	54
CAZ	0	0	0	0	4	13	14	0	0	1	1	5	18	21	16	67
PTZ	0	0	0	0	0	0	0	9	18	40	10	3	2	7	4	17
CIP	0	0	1	2	0	6	2	8	9	1	5	1	26	18	14	97
GEN	0	0	0	0	0	0	7	22	17	10	2	6	2	6	21	51
MER	8	15	39	19	2	2	5	2	1	0	0	0	0	0	0	13
SXT	1	1	16	11	7	4	2	0	7	0	1	43*	0	0	0	55
COL	0	0	0	0	0	0	13	50	26	0	0	4	0	0	0	4

266 \* MIC of 43 isolates >8 SXT

267 Vertical black lines indicate EUCAST ECOFF values. FEP, cefepime; CAZ, ceftazidime; PTZ, piperacillin-  
 268 tazobactam; CIP, ciprofloxacin; GEN, gentamicin; MER, meropenem; SXT, trimethoprim-sulfamethoxazole;  
 269 COL, colistin; NWT, non-wild-type

270

271 **Analysis of resistance profiles between facilities reveal a difference in the proportion of**  
 272 **multidrug resistant *E. coli* isolates**

273 Analysis of resistance was determined using percentages representing the relative proportion  
 274 of isolates characteristic of each facility. This was carried out to mitigate the limitation  
 275 associated with each facility housing a different number of residents. Analysis revealed that  
 276 out of the 58 isolates from Facility 1 that were assessed, 49 (84.5 %) were found to be non-  
 277 susceptible to the fourth-generation cephalosporin, cefepime and 57 (98.3 %) were non-  
 278 susceptible to the third-generation cephalosporin ceftazidime These results contrast with  
 279 those obtained for Facility 2, with only two out of the 27 (7.4 %) *E. coli* isolates displaying

280 low to moderate levels of resistance against both cefepime (MICs of up to 2 mg/L) and  
281 ceftazidime (MICs of up to 16 mg/L) (Figure 1a and Table 1).

282 MIC assays revealed high-level ciprofloxacin resistance of  $\geq 64$  mg/L (more than 1024 times  
283 the ECOFF value) for 34.4 % (n=32/93) of all isolates recovered (Table 1). These isolates  
284 were relatively equally spread between the two facilities (43.1 % and 25.9 % of the total  
285 isolates from Facility 1 and 2 respectively). High-level resistance was also observed against  
286 cefepime and ceftazidime with MICs of  $\geq 64$  mg/L seen in 12.9 % (n=12/93) and 39.8 %  
287 (n=37/93) of all isolates for cefepime and ceftazidime respectively. High-level resistance  
288 against trimethoprim-sulfamethoxazole ( $>8$  mg/L) was observed for 44 of the 93 isolates  
289 (Table 1). Almost all of those isolates (41/58) were from Facility 1 (Figure 1a).

290 Facility 1 harboured a greater number of *E. coli* isolates resistant to multiple antibiotics  
291 (Figure 1b). Facility 2 and Retirement *E. coli* isolates were observed to be primarily  
292 gentamicin and ciprofloxacin resistant. Although all isolates assessed in this study were  
293 resistant to at least one antibiotic, the MDR phenotype was primarily observed in *E. coli*  
294 isolates recovered from Facility 1 (Figure 1b and c), with 54/58 (93.1 %) of these shown  
295 to be resistant to 3 or more antimicrobial classes tested in this study (Figure 1b and c). A  
296 much lower frequency (n=5/27, 18.5 %) of MDR *E. coli* were recovered in Facility 2, which  
297 was observed to harbour mostly non-MDR *E. coli* isolates.

298 The largest disparity observed between the isolates assessed in this study pertained to  
299 gentamicin resistance, with high levels of resistance (MICs of 64-128 mg/L) observed in 66.7  
300 % (n=18/27) of Facility 2 *E. coli* isolates, and only 8.6 % (n=5/58) of Facility 1 isolates.  
301 These findings indicate that initial selection using ciprofloxacin and ceftazidime did not limit  
302 the overall results obtained in this study, and that the prevalence of resistant isolates was not  
303 determined by the number of occupants per facility. Finally, a low incidence of resistance

304 was observed for both meropenem and colistin and was observed in isolates recovered from  
305 Facility 1 only (Figure 1a).

306 To further explore the distribution and association between MDR *E. coli* and isolation site, a  
307 PCA biplot was constructed (Figure 2). As can be seen, the nature and direction of correlation  
308 between resistant isolates and facility indicate that Facility 1 clusters with isolates resistant to  
309 numerous antibiotics – as such displaying an MDR phenotype, whilst Facility 2 clustered  
310 with Retirement isolates displaying resistance to gentamicin and ciprofloxacin only. Unlike  
311 isolates recovered from the two RACFs, which spanned out across the two components,  
312 Retirement isolates clusters tightly and were grouped within the Facility 2 cluster. This  
313 clustering was expected as the resistance profiles for Facility 2 and Retirement isolates were  
314 comparable.

315

#### 316 **Genomic analysis and isolate typing of a subset of AMR *E. coli***

317 Of the 93 wastewater *E. coli* isolates analysed in this study, 35 were selected for WGS.  
318 Selection was based on their antimicrobial resistance profile and time of sample collection  
319 (Table S1). These were selected to further investigate genetic diversity and clonality of the  
320 isolates over the sampling period. As a greater number of MDR *E. coli* were isolated from  
321 Facility 1, these (n=24/35, 68.6 %) made up most of the samples sequenced. The remaining  
322 sequences included seven (20 %) from Facility 2 and four (11.4 %) from Retirement samples.

323 Genotyping and MLST analysis were used here to investigate the genetic diversity of the  
324 isolates sampled in this study. Results revealed 10 different sequence types (ST) and one  
325 unknown sequence type among the 35 sequenced *E. coli* isolates (Figure 3). The pandemic  
326 ST131 was observed in samples isolated from Facility 1 and represented the largest ST  
327 (n=9/24, 37.5 %) of the isolates recovered from this facility. Moreover, two carbapenem

328 resistant *E. coli* isolates that belong to the international high-risk clone ST410 were detected  
329 in Facility 1. Facility 2 was represented by ST1286 (n=4/7, 57.1 %), which was also observed  
330 in Retirement isolates (n=2/4, 50 %). This sequence type is predominantly associated with  
331 environmental and animal isolates [66, 67] and according to the EnteroBase database  
332 (<https://enterobase.warwick.ac.uk/>) has not been isolated previously in Australia. We  
333 observed two *E. coli* isolates of emerging pandemic clone ST1193 in Facility 2. This  
334 sequence type is the most predominant non-ST131 fluoroquinolone resistant ST in the world  
335 [68].

336  
337 To further type the sequenced isolates, phylogroup analysis was carried out based on the  
338 Clermont phylogroup typing method [51] (Figure 4). Of the seven main phylogroups, A, B1,  
339 B2, C, D, E and F [69], 37.1 % (n=13/35) of the sequenced *E. coli* isolates were classified to  
340 the increasingly MDR phylogroup B2. A higher prevalence (n=9/35, 25.7 %) of these isolates  
341 originated from Facility 1. Phylogroup A, representing commensal strains, was the second  
342 most prevalent phylogroup and observed with equal frequencies in Facility 1 and 2 (both  
343 n=5/35, 14.3 %). Meanwhile, Retirement *E. coli* isolates were classified in to phylogroups A  
344 and B only, mirroring Facility 2 isolates. Of the three different sampling sites, Facility 1  
345 isolates represented the greatest diversity in terms of phylogroups, with groups C, D and F  
346 also assigned to these isolates. Strains assigned to group C are closely related to phylogroup  
347 B1 which represent commensal strains [70]. Phylogroup D isolates, like those classified to  
348 group B2 are associated with virulent extra-intestinal *E. coli* infections [71]. Strains assigned  
349 to phylogroup F have been found to be ESBL-producing and resistant to fluoroquinolones  
350 [72].

351 Finally, core genome SNP analysis was used to assess clonal relatedness of all 35 sequenced  
352 *E. coli* isolates, with the results represented in a maximum likelihood phylogenetic tree

353 (Figure 5). Isolates were aligned to *E. coli* EC958 (accession number: NZ\_HG941718.1),  
354 which represents an international MDR strain. Analysis revealed a clustering of isolates by  
355 facility, sequence type, susceptibility profile and phylogroup indicating clonality.  
356 Nonetheless, variability between AMR determinants and plasmid replicons was also observed  
357 between clonally related isolates.



## 358 Identification of *E. coli* AMR genes

359 Given the observations that the wastewater isolates assessed in this study displayed a similar  
360 resistance profile per facility, we sought to evaluate their genotypic diversity and examined  
361 their genomes for the presence of mobile AMR genes and potential plasmids. All these  
362 wastewater isolates were shown to possess a plethora of AMR genes, contributing both to  
363 their intrinsic and acquired resistance. Twenty-seven different AMR genes conferring  
364 resistance to the compounds assessed in this study, which included cephalosporins,  
365 carbapenems,  $\beta$ -lactam-  $\beta$ -lactam- inhibitor combinations, fluoroquinolones,  
366 aminoglycosides, trimethoprim, and sulphonamides, were identified (Figure 5). The extended  
367 spectrum beta-lactamases (ESBLs) such as the *bla*<sub>CTX-M</sub> type genes accounted for 37.3 %  
368 (n=19/51) of the beta-lactamase genes detected in the wastewater samples. The *bla*<sub>CTX-M</sub> type  
369 genes (n=19) included *bla*<sub>CTX-M-15</sub> (n=4/19, 21.1%), *bla*<sub>CTX-M-27</sub> (n=9/19, 47.4 %) and *bla*<sub>CTX-</sub>  
370 *M-62* (n=6/19, 31.6 %). Except for one *bla*<sub>CTX-M-62</sub> carrying strain isolated from Facility 2, all  
371 other *bla*<sub>CTX-M</sub> type genes were found in Facility 1 isolates. Non- ESBL beta-lactamases,  
372 *bla*<sub>OXA-1</sub> (n=4/35, 11.4 %) and *bla*<sub>OXA-181</sub> (n= 2/35, 5.7 %) were identified in isolates  
373 recovered from Facility 1 only, whereas *bla*<sub>TEM-1</sub> was detected in *E. coli* isolated from Facility  
374 1 (n=8/35, 22.9 %), Facility 2 (n=5/35, 14.3 %) and retirement (n=2/35, 5.7 %). No  
375 carbapenemase genes were detected despite meropenem resistance in 4 out of the 6  
376 sequenced meropenem resistant *E. coli* isolates (Figure 5). However, these isolates harboured  
377 at least one ESBL or AmpC beta-lactamase genes. Non-carbapenemase-producing low level  
378 carbapenem-resistance in *E. coli* could be mediated by ESBL or AmpC beta-lactamases  
379 associated with an overexpression of efflux pumps (such as AcrAB) or a loss of porin  
380 (OmpF) expression [73, 74].

381 Bacteria with ESBL phenotypes are frequently found to harbour additional resistance genes.  
382 In this study, isolates were found to carry the plasmid-mediated quinolone resistance

383 (PMQR) genes, which were detected in Facility 1 isolated strains only (n=14/35, 40 %). Of  
384 these, the most prevalent PMQR gene detected was *qnrB4* (n=6/14, 42.9 %). Also only  
385 isolated from Facility 1 was the ciprofloxacin and aminoglycoside modifying enzyme (*aac-*  
386 *(6')-Ib-cr*), accounting for 14.3 % (n=5/35) of the total isolates detected. Genes mediating  
387 aminoglycoside resistance such as *aac(3)-IId* were also detected in 7 out of 35 isolates (20 %)  
388 and exclusively recovered from Facility 2 and Retirement, whereas *aph3Ib* and *aph6Id* (both  
389 n=13/35, 37.1 %), which are commonly found on plasmids, integrative elements and  
390 chromosomal islands, were detected in wastewater isolates recovered from Facility 1. A  
391 higher prevalence of the sulfonamide genes, *sul1* (n=21/35, 60 %) and *sul2* (n=15/35, 42.9  
392 %) and trimethoprim *dfrA17* (n=15/35, 42.9 %) resistance genes were also observed, with the  
393 majority detected in Facility 1 isolates. Lastly, ciprofloxacin resistance due to point mutations  
394 was also investigated, with 32/35 (91.4 %) isolates shown to be phenotypically ciprofloxacin  
395 resistant. Of these, 31/32 (96.9 %) were shown to possess mutations within *gyrA*, *parC* or  
396 *parE* genes (Table S2).

397

### 398 **Prevalence of plasmid replicon types detected in AMR *E. coli* isolates**

399 Plasmids are known to carry and distribute AMR genes such as ESBLs, carbapenemases, and  
400 PMQR genes [75]. Due to their ability to transmit AMR genes to other bacteria, they can  
401 enable the rapid evolution of bacteria in response to different environmental pressures [76].  
402 In this study, analysis of plasmids within the 35 sequenced wastewater *E. coli* isolates were  
403 differentiated based on plasmid replicon types. The incompatibility groups (Inc) IncFIB  
404 (n=22/35, 62.9 %) and IncFII (n=22/35, 62.9 %) accounted for the prevalent replicon types  
405 detected in the *E. coli* isolates. All isolates were positive for at least one Inc or replicate (Rep)  
406 group. Analysis of the most dominant plasmid replicon type by facility revealed Facility 1  
407 isolates to harbour IncF (IncFIA, IncFIB, IncFIC and IncFII) type plasmids, Facility 2 Col

408 (ColBS512, ColMG828, ColMP18, Col440I and Col156) type plasmids, with Retirement  
409 isolates harbouring both types with near equal distribution.

410

#### 411 **Relative abundance of bacterial families represented by metagenomics**

412 To compare the complete wastewater resistome with data obtained from the *E. coli* isolates,  
413 metagenomics sequences were obtained from all sample sites and used here to assess  
414 taxonomy and prevalence of AMR in each site. Taxonomic annotation of the metagenomic  
415 datasets revealed the 25 most abundant genera detected in Facility 1, 2 and Retirement  
416 wastewater samples (Figure 6). Bacteria belonging to the *Escherichia* genus represented an  
417 average of 3.6 % of bacteria in Facility 1, 0.7 % Facility 2 and 0.9 % in Retirement. A  
418 comparison between Facility 2 and Retirement revealed a greater number of AMR *E. coli*  
419 isolates recovered from Facility 2 despite Retirement presenting an overall greater abundance  
420 of these organisms. This exemplifies the importance of culturing in the assessment of  
421 prevalence of AMR or MDR organisms within wastewater samples.

422

#### 423 **Occurrence and abundance of AMR genes in wastewater samples**

424 Given the phenotypic-genotypic relationship in *E. coli* [77-79], this study focused on  
425 acquired resistance genes, which, as a result of horizontal gene exchange, have rendered  
426 many important pathogens such as MDR *E. coli*, potentially unresponsive to current  
427 treatment. The focus of this study included the ESBL *bla*<sub>CTX-M</sub>, beta-lactamases *bla*<sub>OXA</sub>,  
428 *bla*<sub>TEM</sub> and *bla*<sub>IMP</sub>, the dihydrofolate reductase genes *dfrA1* and the dihydropteroate synthases  
429 *sul1* and *sul2*, which confer bacterial resistance against trimethoprim and sulphonamide. In  
430 addition to these, PMQR genes *qnrB* and *qnrS* genes and *aac(6')-Ib-cr* gene (cr for

431 ciprofloxacin resistance) were also evaluated. Finally, prevalence of gentamicin resistance  
432 was explored by the assessment of the *aac(3)*- and *aac(6')* – genes.

433 A higher abundance of most of the acquired resistance genes assessed in this study, by a  
434 comparison of TPM values obtained for the AMR genes, were detected in samples  
435 originating from Facility 1 (Figure 7a). The exception to this were Facility 2 samples which  
436 had a larger proportion of the gentamicin resistance genes *aac(3)*- and *aac(6')*, and  
437 Retirement samples which had a higher prevalence of the beta-lactamase *bla<sub>IMP</sub>*. The  
438 differences between Facility 1 and 2 samples regarding the number of the different AMR  
439 genes assessed here were determined to be significant ( $p < 0.05$ ; Figure 7b). Differences were  
440 seen in beta-lactamase *bla<sub>CTX-M</sub>*, *bla<sub>OXA</sub>* genes, the dihydrofolate reductase *dhfrA1* genes, the  
441 3-N-aminoglycoside acetyltransferase *aac(3)* and the 6'-N-aminoglycoside acetyltransferase  
442 *aac(6')* genes, as well as the sulfonamide *sul1* and *sul2* genes and the quinolone antibiotic  
443 resistance genes *qnrS*. These results clearly show that Facility 2 has a lower relative  
444 abundance of the mobile AMR genes assessed in this study, all of which have been associated  
445 with human pathogens. No significant differences were found for the AMR abundance  
446 between Facility 2 and Retirement samples.

447

## 448 **DISCUSSION**

449 The rise of AMR and emergence of bacteria that are resistant to multiple classes of antibiotics  
450 presents a global threat to human health [80]. Measures taken to fight resistance include  
451 broadening our understanding of the selective pressures driving AMR to be able to mitigate  
452 them. RACFs have been identified as sites harbouring elevated levels of AMR bacteria and  
453 presenting an environment which is highly selective to their development and emergence [81,  
454 82].

455 In this study, the prevalence of AMR *E. coli* was used as an indicator of the incidence and  
456 persistence of bacterial resistance in two RACFs, one of which has implemented an AMS  
457 program (Facility 2), whilst the other has not (Facility 1). Assessment of the prevalence of *E.*  
458 *coli* in these two RACFs and in one Retirement village was carried out with an in-depth  
459 analysis of 93 AMR *E. coli* isolates recovered from wastewater samples, collected over five  
460 different time points. Results revealed a high proportion of MDR *E. coli* isolates (66.7 %)  
461 which were more frequently recovered from Facility 1. Isolates recovered from Facility 1  
462 displayed high levels of resistance against ceftazidime, cefepime and ciprofloxacin with MIC  
463 values of >128, >512 and >1024 times above the ECOFF. While Facility 2 isolates displayed  
464 high-level gentamicin resistance, with MIC values of >64 mg/L (>32 times above the  
465 ECOFF), a trend not observed in Facility 1 *E. coli* isolates. These results highlight several  
466 concerns, including the high prevalence of ESBL producing *E. coli* in Facility 1 and the  
467 number of ciprofloxacin resistant *E. coli* isolates prevalent in both facilities.

468 According to the Antimicrobial Use and Resistance in Australia (AURA) Surveillance  
469 System (2019) rates of resistance for many priority organisms, including *E. coli*, have not  
470 changed significantly over the last 2 years. However, there are several exceptions to this,  
471 including increasing rates of resistance observed in *E. coli* against commonly used agents  
472 such as ceftriaxone, a third-generation cephalosporin and ciprofloxacin, amongst other  
473 fluoroquinolones, despite restricted access in the community and guidelines advising these  
474 not to be prescribed when other options are available, and if needed to be prescribed for the  
475 shortest duration clinically possible (AURA, 2019). Of the two cephalosporins assessed here,  
476 higher rates (66.7 %) of resistance were seen for the third-generation ceftazidime, than for the  
477 fourth-generation extended spectrum agent, cefepime (53.7% of all isolates). Whole genome  
478 sequencing of ESBL producing *E. coli* revealed the presence of CTX-M type genes assigned  
479 to ST131 isolates. This sequence type is a major contributor to both sepsis and urinary tract

480 infections (UTIs) in Australia [83]. Moreover, phylo-typing assigned these isolates to the B2  
481 phylogenetic group. This group is frequently associated with CTX-M-15 producing *E. coli*  
482 [84], although in this study was associated with CTX-M-27 producing isolates. Globally,  
483 CTX-M-15 producing strains are recognised as the main drivers of the pandemic spread of  
484 this sequence type [85]. However, genomic comparisons of ST131 isolates have shown that  
485 both CTX-M-15 and CTX-M-27 producing isolates dominate in Australia [83]. CTX-M-27  
486 ESBLs were also initially postulated to be confined to ST131 [86], however, these have now  
487 also been identified in ST38 *E. coli* isolates, as also observed in this study, suggesting an  
488 emergence of new ESBL-producing clones [87]. The dissemination and continual emergence  
489 of these ESBLs and others are facilitated by their presence on plasmids belonging to the IncF  
490 group [88], amongst others, which were also detected here. This plasmid group is often  
491 associated with *bla*<sub>CTX-M-15</sub>, *bla*<sub>TEM-1</sub>, *bla*<sub>OXA-1</sub> and *aac(6')lb-cr* resistance genes, all of which  
492 were detected here and largely in isolates recovered from Facility 1.

493 The isolation of ST131 in this study highlights the continual spread of this pandemic clonal  
494 group. Although a worrying trend, the presence of ST131 *E. coli* isolates within RACFs has  
495 been observed worldwide [89, 90]. It has been postulated that isolates of this sequence type  
496 have superior transmission and colonisation abilities, allowing for rapid global proliferation  
497 [91-93]. These characteristics coupled with the intrinsic nature of RACFs, and residents  
498 therein makes the presence of this clonal group within RACFs a significant concern.

499 In addition to being resistant to extended spectrum cephalosporins, ST131 *E. coli* clones are  
500 also associated with fluoroquinolone resistance [94] which can be coupled with  
501 aminoglycoside and trimethoprim-sulfamethoxazole resistance [92]. In this study, all ST131  
502 isolates were ciprofloxacin resistant and except for one isolate, trimethoprim-  
503 sulfamethoxazole resistant. Worryingly, two isolates were extensively drug resistant with  
504 resistance against all the tested classes of antibiotics observed. Fluoroquinolone resistance

505 continues to be of concern and was observed in this study with high levels of ciprofloxacin  
506 resistance (MICs of > 64 mg/L) detected in isolates recovered from both facilities, and in  
507 Retirement samples. According to the Australian Group of Antimicrobial Resistance (AGAR)  
508 Sepsis Outcome Programs, 2021, rates of fluoroquinolone resistant *E. coli* have continued to  
509 increase from 13.7 % in 2013 to 21.8 % in 2021. Increasing trends have also been observed  
510 in Europe with the European Antimicrobial Resistance Surveillance Network (EARS-Net)  
511 reporting 25.3 % fluoroquinolone resistant *E. coli* isolated in 2018, representing an increase  
512 from previous years [95]. This upward trend is also reported in *E. coli* isolated in Asia as well  
513 as North America [96]. Various theories have been presented regarding the acquisition of  
514 fluoroquinolone resistant *E. coli* including possible colonisation with these as a result of  
515 hospital admissions [97], the continual use of these in treatment of UTIs [98] and clonal  
516 spread [99].

517 Analysis of fluoroquinolone resistance genes within WGS isolates revealed the presence of  
518 numerous ciprofloxacin resistance genes including the transferable plasmid-mediated  
519 quinolone resistance (PMQR) determinants *qnr* and *aac(6)-Ib-cr*. The presence of these is of  
520 concern as these are able to be transferred to other bacteria, are often associated with  
521 elements such as insertion sequences, transposons, phages and plasmids which enhances their  
522 transferral and can be accompanied by other mobile resistance genes [100].

523 Metagenomic analysis of the prevalence of these and other mobile resistance genes within the  
524 sampled wastewater revealed that Facility 1 harboured a greater number of mobile resistance  
525 determinants. A limitation of using the metagenomic method for the assessment of bacterial  
526 and AMR gene abundance is that each sample of wastewater represents a single point in time.  
527 As such, wastewater components may be influenced by the amount of faecal matter in the  
528 samples, the activities being performed at the time of collection, such as cleaning or washing  
529 and the detergents being used, all of which may alter the microbial and antimicrobial gene

530 abundance within each sample. To minimise this limitation, sites were sampled over a 10-  
531 hour period at five different time points throughout an 18-month period and samples assessed  
532 by microbial culturing and metagenomics. To ensure that the data are not influenced by the  
533 number of people, flow rate at the timepoints of grab sample collection or any other potential  
534 variables, the data obtained from metagenomics was normalised and AMR gene abundance  
535 was calculated as transcripts per million mapped bacterial reads. In addition to Facility 1  
536 samples, which displayed consistently higher levels of bacteria belonging to the *Escherichia*  
537 genus, metagenomic results also revealed a higher prevalence of these in Retirement  
538 wastewater samples when compared to Facility 2, but an overall lower prevalence of AMR *E.*  
539 *coli* upon culturing. As such, the higher prevalence of AMR *E. coli*, as identified by  
540 microbial culturing and metagenomics for all five sampling periods, is indicative of a general  
541 trend only seen for Facility 1. This trend was also observed with respect to mobile resistance  
542 genes, with Facility 1 displaying a comparatively higher abundance of certain resistance  
543 genes.

544 Analysis of the mobile resistance genes revealed plasmid-mediated resistance genes, some of  
545 which are associated with Class 1 integrons that have been isolated from AMR clinical  
546 isolates belonging to both the Enterobacteriaceae and Pseudomonaceae families [101].  
547 Having the ability to integrate gene cassettes which confer resistance to a broad range of  
548 antibiotics has made the prevalence of Class 1 integrons a proxy for bacterial resistance  
549 [102]. As such, the greater presence of the mobile resistance genes may be indicative of a  
550 higher presence of AMR bacteria in Facility 1 wastewater.

551 Limitations of this study include that only two RACFs were assessed and compared, and that  
552 these differed in the number of residents residing in them. However, all variables that were  
553 reported in this study were the same between the facilities, for example the age of the cohort  
554 and type of care provided. Moreover, the two facilities were managed by the same age care



555 provider, which runs additional facilities in Australia. It is possible that there could have been  
556 differences in variables that were not measured as part of this study such as the socio-  
557 economic status of the demographic, duration of stay, staff turnover number or age of the  
558 facility. To overcome variations in resident numbers and isolate prevalence, percentages were  
559 utilized for standardized comparisons, allowing meaningful assessments across facilities in  
560 the study. However, to our knowledge this is the first study to include such a comprehensive,  
561 all-encompassing testing protocol. This study has included isolation of *E. coli* and culture  
562 based genotypic AMR analysis, genomic analysis of the resistome of individual organisms  
563 and metagenomic sequencing and analysis of the total wastewater resistome. It is also the  
564 first time that a direct comparison of two RACFs has been undertaken, with a spotlight on  
565 AMR.

566 The findings of this study offer an insight into bacterial resistance and the prevalence thereof  
567 within RACFs. Here, one facility was found to harbour higher rates of AMR *E. coli*,  
568 including the pandemic clone ST131. Moreover, metagenomic analysis revealed a higher  
569 incidence of mobile resistance determinants in wastewater sourced from this facility. The  
570 inclusion of microbial culturing in combination with whole genome sequencing and  
571 metagenomic analysis has allowed for a more complete assessment of the RACFs evaluated  
572 in this study. Based on the differences in the prevalence of AMR *E. coli* and mobile  
573 resistance genes within the RACFs assessed here, our results show that each facility presents  
574 a unique environment which may select for resistance and possible transmission of AMR  
575 bacteria. Mitigating this may be the implementation of an AMS program, which was present  
576 in Facility 2. However, it is understandable that the implementation of AMS programs in  
577 RACFs can be complicated and is dependent on numerous factors including access to  
578 resources, staff education and a high prevalence of bacterial colonisation and challenges in

579 diagnosing and treating infections [103]. Nonetheless, the benefits of such programs can help  
580 mitigate the spread and continual development of AMR.

581 Currently there are significant gaps in the surveillance of AMR and MDR organisms within  
582 RACFs. Our understanding of the colonisation of such organisms in RACF residents is also  
583 lacking, as is knowledge pertaining to the risk factors which aid both the development of  
584 AMR and potential spread of both AMR and MDR organisms and high-risk antimicrobial  
585 genes from RACFs to the wider community. The results generated in this study highlight the  
586 need for ongoing surveillance studies focussing on RACFs. Which, given our ageing  
587 population, are greatly needed to address and possibly mitigate the ongoing threat of AMR.

588

#### 589 **SUPPLEMENTARY MATERIAL**

590 Supplemental 1. AMS Strategies

591 Table S1. Overview of antibiotic resistance profiles, with MIC and gene profiles (0 = absent  
592 or 1 = present) of wastewater *E. coli* isolated from Facility 1 and 2 and from Retirement,  
593 Adelaide, Australia.

594 Table S2: GyrA, ParC and ParE mutations in 35 wastewater *E. coli* isolates corresponding  
595 ciprofloxacin MICs.

596

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610

611 **Figure 1:** Facility 1 shows the highest prevalence of MDR *E. coli* isolates. **(A)** Percentage of total  
612 AMR *E. coli* isolates per facility and retirement village. **(B)** Prevalence of *E. coli* isolates (n=93) per  
613 facility and number of antibiotics (total n=8) each isolate was resistant against. **(C)** Number of MDR  
614 and non-MDR *E. coli* isolates per facility. Facility 1 n=58, Facility 2 n=27 and Retirement n=8. FEP,  
615 cefepime; CAZ, ceftazidime; PIP-TAZ, piperacillin-tazobactam; CIP, ciprofloxacin; GEN,  
616 gentamicin; MER, meropenem; SXT, trimethoprim-sulfamethoxazole; COL, colistin, Ab (s),  
617 antibiotic(s).

618

619 **Figure 2:** Principal component analysis (PCA) plot shows clustering of MDR *E. coli* isolates  
620 recovered from Facility 1 (blue) relative to isolates which are resistant to gentamicin (GEN) and  
621 ciprofloxacin (CIP) only and recovered from Facility 2 (red) and Independent Living Facility  
622 (Retirement, grey). Ellipses were drawn at a confidence level of 0.95. CEP, cefepime; CTZ,  
623 ceftazidime; PIT, piperacillin-tazobactam; CIP, ciprofloxacin; GEN, gentamicin; MER, meropenem;  
624 TRS, trimethoprim-sulfamethoxazole; COL, colistin.

625

626 **Figure 3.** Distribution and percentage of sequenced *E. coli* isolate MLSTs per facility and retirement  
627 village. Facility 1 n=24, Facility 2 n=7 and Retirement n=4.

628

629 **Figure 4:** Phylogroup distribution of 35 *E. coli* isolates from two RACFs and a Retirement site.  
630 Facility 1 n=24, Facility 2 n=7 and Retirement n=4.

631

632 **Figure 5:** Phylogenetic SNP analysis of 35 AMR *E. coli* isolated from RACFs and a Retirement  
633 Village in South Australia, corresponding resistance profile presence of AMR determinants and  
634 plasmid replicons. SNP analysis was performed using the isolate EC958 (accession number:  
635 NZ\_HG941718.1) as the reference genome. Vertical columns demonstrate: (1) isolation site, (2)  
636 phylogroups (3) resistance phenotype: resistant (filled squares) and sensitive (empty squares) (4)  
637 resistance genotype including plasmid mediated quinolone resistance (PMQR): presence (filled  
638 circles) and absence (empty circles) of acquired resistance genes (5) mutations in the quinolone  
639 resistance-determining regions: present (blue filled rectangle) and absent (empty rectangle) (7)  
640 predominant plasmid replicon types: presence (filled left side triangles) and absence (empty left side  
641 triangles). FEP, cefepime; CAZ, ceftazidime; PIT, piperacillin-tazobactam; CIP, ciprofloxacin; GEN,  
642 gentamicin; MER, meropenem; SXT, trimethoprim-sulfamethoxazole; COL, colistin.

643

644 **Figure 6.** Relative abundance (%) of the top 25 genera identified in wastewater collected from two  
645 RACFs and Retirement samples. Red bars represent bacteria belonging to the *Escherichia* genus.

646

647 **Figure 7:** Distribution of mobile resistance genes in wastewater from RACFs and an independent  
648 living facility (Retirement). (A) Heatmap of distribution and abundance clustering of resistance genes.  
649 AMR genes are shown on the right vertical axis. Hierarchical clustering was carried out on both  
650 resistance genes and sites. AMR abundance is normalised as TMP (transcripts per million reads). (B)  
651 Boxplots of select AMR gene reads per ten million (TPM) per facility. Each sample is represented by  
652 a dot with horizontal jitter for visibility. The horizontal box lines represent the first quartile, the  
653 median, and the third quartile. Whiskers represent the range of points within the first quartile – 1.5×  
654 the interquartile range and the third quartile + 1.5× the interquartile range. Differences were  
655 considered statistically significant at \*  $p < 0.05$ ; \*\* $p < 0.01$  level, determined by a two-tailed Mann-  
656 Whitney U test.

657

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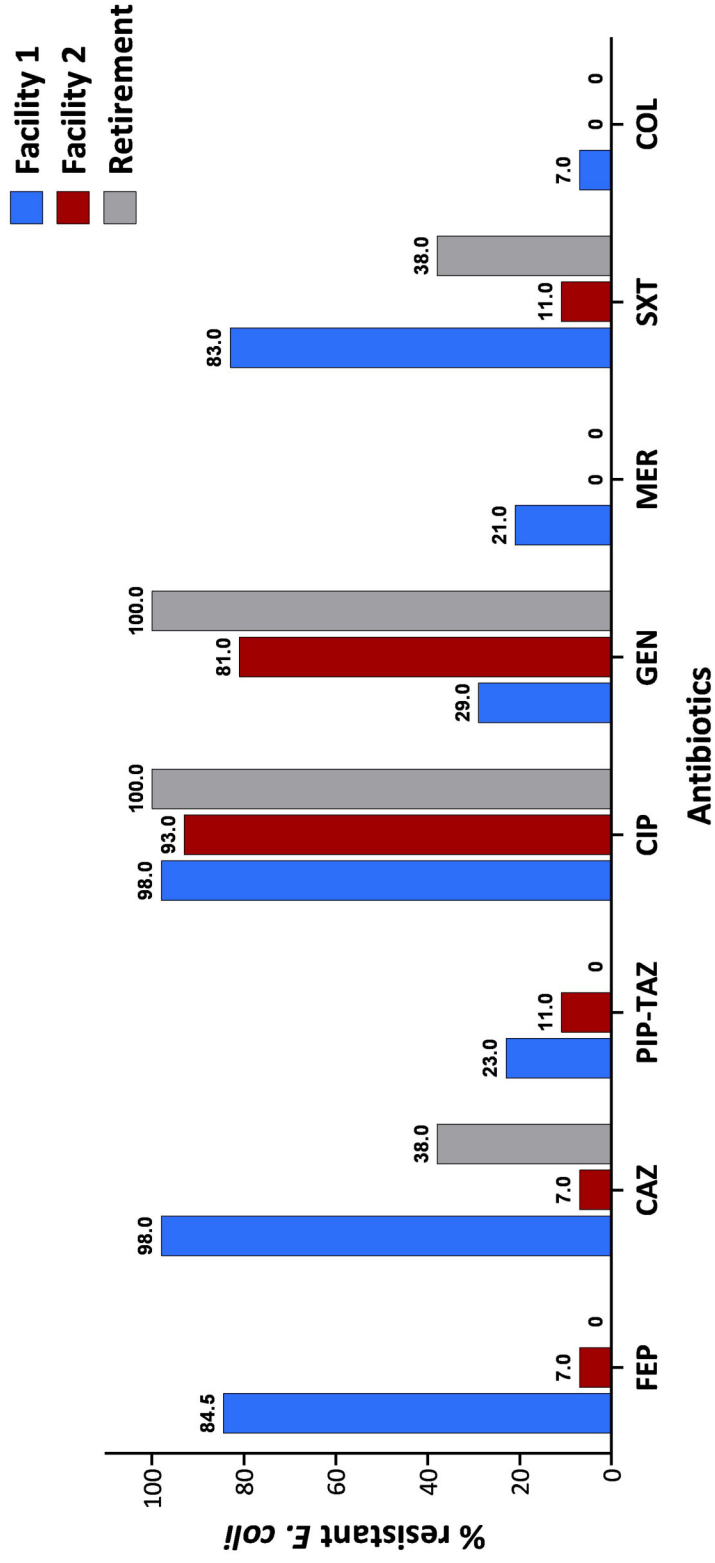
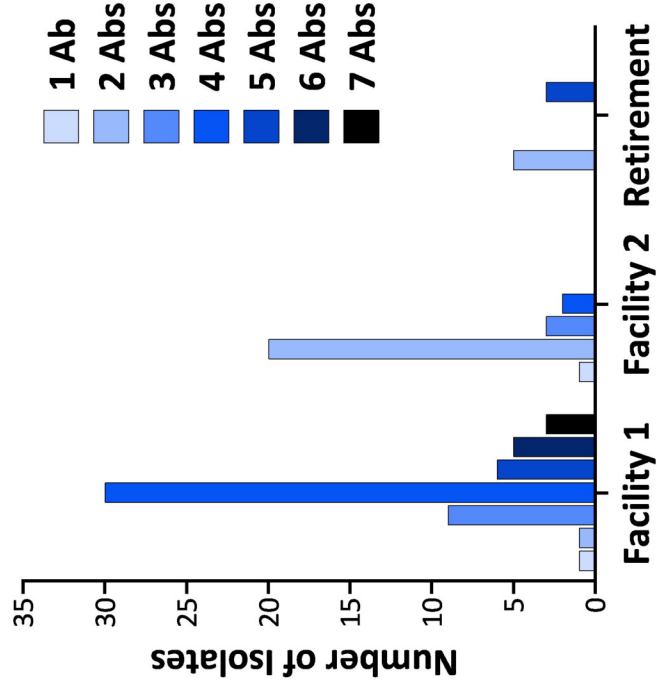
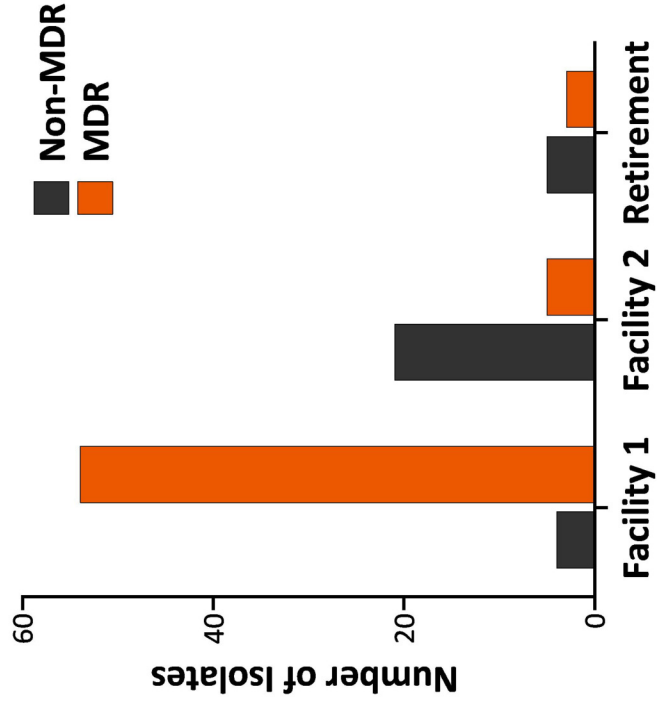
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### PCA - Biplot

